

# **Physiological condition, short term survival and predator avoidance behaviour of discarded Norway lobsters (*Nephrops norvegicus*)**

**Amaya Albalat<sup>1\*</sup>, Arnaud Collard<sup>1</sup>, Bruce McAdam<sup>1</sup>, Christopher J. Coates<sup>2</sup>, Clive J. Fox<sup>3</sup>**

<sup>1</sup>*Institute of Aquaculture, School of Natural Sciences, University of Stirling, Stirling FK9 4LA, UK*

<sup>2</sup>*Department of Biosciences, College of Science, Swansea University, Swansea SA2 8PP, UK*

<sup>3</sup>*Scottish Association for Marine Science, Dunstaffnage, Oban PA37 1QA, UK*

*\*Corresponding author: tel: +44 (0)1786 467917; e-mail: amaya.albalat@stir.ac.uk*

**Accepted for publication in *Journal of Shellfish Research*, published by National Shellfisheries Association.**

The landing obligation of the reformed European Union Common Fisheries Policy is designed to encourage more selective fishing strategies and improve recording of catches. There are allowable exemptions to this landing obligation including for species with high post-release survivability. Discarding patterns of prawns (*Nephrops norvegicus*) were evaluated in a trawl fishery in the Firth of Clyde, West Scotland which supplies the live-catch market. Around 30% of the *Nephrops* caught were discarded but the reasons for discarding changed seasonally. Using visual indices, physiological biomarkers and video recordings this study evaluated the physiological condition linked to short-term survival and predator avoidance behaviour of the discarded animals. Although short-term survival after 48 h recovery was high (around 90%) and physiological measures indicated that discarded *Nephrops* can recover from trawling, survival was negatively affected by levels of physical damage and *Hematodinium* infection. Taking into consideration these factors a conservative estimate for discard survival was 63-88%. Underwater video showed that *Nephrops* discarded in good condition rapidly recovered normal behaviour when placed on the seabed. Moribund

animals, however, took up to 10 minutes to return to an upright posture and this time was sufficient for predators to be attracted. Since around 20% of *Nephrops* were in a moribund condition immediately after trawling, the survival estimates based upon enclosed recovery experiments may need correcting by up to this amount to account for potential interactions with predators on the seabed. The post-release survival rates in discarded *Nephrops* suggested for this fishery are relatively high compared with other *Nephrops* trawl-fisheries which have been studied. This could be explained because this fishery targets the live-market, prioritises product quality over volume and uses short-duration tows leading to relatively low levels of physical damage to the *Nephrops*.

1

2 **Key words:** discards, physiological condition, behaviour, *Nephrops*

3

4

## 5 1. Introduction

6 Discards are target or non-target animals that are not retained on-board commercial  
7 fishing vessels but are returned to the sea, either alive or dead. Organisms may be discarded  
8 for a variety of reasons including lack of fisheries quota, lack of demand or where the  
9 organisms are unmarketable due to other reasons such as size, damage or disease (Catchpole  
10 et al., 2005a,b). The problem of discarding in European fisheries has been widely discussed  
11 from ecological, economic, ethical and environmental perspectives (Eliassen et al., 2014) and  
12 received increasing public attention in the run-up to the 2013 reforms of the European  
13 Union's Common Fisheries Policy (Borges, 2015). As a result a 'landing obligation' was  
14 introduced and is being sequentially applied across EU fisheries. The aims of this policy  
15 change are to reduce waste, encourage uptake of more selective fishing practices and to  
16 improve data recording of actual catches. Consequences on fisheries management of the  
17 recent United Kingdom vote to exit the EU are unclear but it is possible the 'landing  
18 obligation' will be retained as various forms of discard bans are applied in other non-EU  
19 countries such as Norway and Iceland.

20 Under the CFP landing obligation discarding can potentially continue where there is  
21 evidence that the species has a high rate of post-discard survival. Forcing landing of these  
22 discards, and their subsequent disposal, would only increase economic costs for little overall  
23 conservation benefit. The prawn *Nephrops norvegicus* has been identified as having  
24 potentially high post-release survival and a landing obligation exemption has been approved  
25 by the European Union for creel-caught *Nephrops* (Ungfors et al., 2013; Fox, 2014; Marine  
26 Scotland, 2015). On the other hand, scientific studies suggest that trawl-caught *Nephrops* do  
27 not necessarily survive as well as creel-caught animals. Despite having a hardened  
28 exoskeleton, *Nephrops* seem to be quite vulnerable to physical damage, particularly puncture

29 wounds to the carapace, which can lead to a loss of haemolymph and eventual mortality  
30 (Wileman et al., 1999). Levels of physical damage will be affected by factors such as gear  
31 type, volume and composition of the catch, trawling time, locations of animals within the  
32 cod-end, the care with which the catch is hauled and handled on board as well as the size and  
33 condition of the animals (Milligan et al., 2009). The eyes of *Nephrops* are also particularly  
34 vulnerable to light exposure at levels typically found at the sea surface, although there is no  
35 evidence linking such damage with subsequent reduced survival (Gaten et al., 2013). From a  
36 physiological perspective, the process of being caught in a trawl is stressful with anaerobic  
37 metabolism being activated in an attempt to maintain energy levels (Ridgway et al., 2006b;  
38 Albalat et al., 2009). Nevertheless, *Nephrops* do show considerable ability to recover from  
39 such stress if post-harvest practices are optimised and aerial exposure is minimised (Albalat  
40 et al., 2010). In such situations key physiological parameters, such as the replenishment of  
41 muscle ATP, have been reported to recover to un-stressed levels within 6 h post-capture with  
42 other anaerobic metabolites (i.e. L-lactate and muscle glycogen) returning to rested conditions  
43 within 24-48 h (Albalat et al., 2010). The potential of *Nephrops* to recover from trawling is  
44 also influenced by season (Lund et al., 2009; Albalat et al., 2010) although it is not clear if  
45 this is mainly due to temperature or other seasonally varying factors, such as animal size, sex  
46 and moult status (Milligan et al., 2009).

47       Most post-release survival studies conducted to date have monitored the recovery of  
48 animals in tubes or cages, for example Harris and Andrews (2005). These approaches provide  
49 estimates of physiological recovery but because the recovering *Nephrops* are protected from  
50 predation they may under-estimate true post-discard mortality (STECF, 2015). In behavioural  
51 terms survival will be related to (i) the length of time taken to regain normal predator  
52 avoidance behaviour (ii) the abundance of potential predators in the immediate area (iii) the  
53 ability of the animals to escape predators using their tail-flip response and (iv) whether the

54 animals are discarded over suitable muddy habitat allowing them to find burrows in which to  
55 hide.

56 In the Firth of Clyde there are around 18 registered under 10 m *Nephrops* trawlers the  
57 majority of which fish locally (Fox et al., 2015) and supply the live-market. Fishing is  
58 typified by relatively short tows (<2.5 hours) using single-rig trawls fishing relatively close to  
59 port. The *Nephrops* are generally sorted from the catches within 60-90 mins of hauling, and  
60 placed into boxes of tubes according to size (Supplementary Fig. 1). The tubed animals are  
61 then held in seawater tanks on-board before being transferred to shore-based tanks where  
62 they are stored for up to 3 days before being shipped by truck, mainly to continental Europe.  
63 Prawns unfit for market are discarded during sorting over muddy grounds as the vessels  
64 prepare for the next tow. Using a commercial day vessel operating in this fishery this study  
65 analysed 1) composition of retained and discarded *Nephrops* 2) the physiological condition  
66 and short-term survival of captive *Nephrops* 3) predator avoidance behaviour using  
67 underwater video and 4) which variables are linked to survival.

## 68 **2. Materials and Methods**

### 69 *2.1 Trawling and data collection*

70 Trials were conducted in the Firth of Clyde, Scotland using a commercial vessel  
71 (Eilidh Anne GK2) which targets *Nephrops* for the live market. Fishing was undertaken using  
72 a single-rig Harkess (Prestonpans, Scotland), rock-hopper trawl fitted with an 85 mm  
73 diamond mesh net and a square-mesh 120 mm Cod Recovery Zone Panel (standard  
74 commercial net and rig). Three sets of *Nephrops* recovery trials were conducted in February,  
75 March and June and a behavioural trial in July 2015. Environmental conditions and trawling  
76 coordinates were recorded for each tow (Supplementary Table 1). Temperature and salinity

77 profiles were recorded in February and March using an YSI Castaway CTD. In the June trial  
78 surface temperature readings were obtained from the vessel but no salinity measures could be  
79 taken due to instrument failure.

80 The skipper was asked to follow his normal fishing practices and once the catch was  
81 on board he began sorting the *Nephrops* into four commercial size categories: Large,  
82 medium, small and extra-small. In this fishery animals in the first three categories are  
83 destined for the live market while the extra-small *Nephrops* are marketed as whole fresh or  
84 frozen langoustines. Animals below the Minimum Conservation Reference Size (carapace  
85 length <20 mm) or not fit for market according to the skipper's criteria were placed in a  
86 separate basket as 'discards'. After sorting of each tow was completed animals in the large,  
87 medium and small categories were counted and their carapace lengths recorded using digital  
88 calipers. Because of the numbers of animals in the extra-small and discarded grades, size  
89 profiles were based on measuring animals within a weighed sub-sample, which was then  
90 raised by the total catch weight of that grade.

## 91 *2.2 Physiological condition and short-term survival of captive discarded Nephrops*

92 Fifteen *Nephrops* were taken from the discards category as soon as practical after the  
93 skipper had begun sorting the catch and sampled to provide baseline physical and  
94 physiological information (time = 0). This generally took place within 15 minutes of the  
95 catch being hauled aboard. For each individual, carapace length (CL), sex, damage, damage  
96 (described below), vigour index (described below), moulting stage, the presence of mature  
97 gonads in females (green gonads) and any obvious *Hematodinium* symptoms were recorded.

98 Visible damage was scored against the three-levels (0=no damage, 1=slight damage,  
99 2=severe damage) as introduced by Ridgway et al. (2006a). Injuries were not counted if there  
100 was evidence of tissue regeneration, indicative of an old injury. Vigour was assigned to five  
101 categories: A, B, C, moribund or dead according to the posture of the animal when held mid-

102 air, the position of the antennae and claws, the degree of tail tension indicative of reflex  
103 impairment and movement of walking legs (Supplementary Table 2). Infection by  
104 *Hematodinium* was evaluated based on their body colour (Field et al., 1992) as symptomatic  
105 of patent infection or as asymptomatic (which includes both truly uninfected and also sub-  
106 patently infected individuals). The validity of these visual criteria for indicating patent  
107 infection has recently been confirmed using molecular methods for detecting and quantifying  
108 the parasites in the haemolymph (Beevers et al., 2012). Carapace hardness was used as a  
109 straightforward measure of the moulting stage of individual *Nephrops*. Animals were  
110 classified as ‘soft’ if gentle squeezing just behind the eyes gave a clear distortion (Milligan et  
111 al., 2009). ‘Soft’ animals included those recently moulted and at late inter-moult (i.e. with  
112 calcium withdrawn from exoskeleton). For consistency visual indices (vigour, damage, body  
113 colour and carapace hardness) were recorded by same observers in all trips.

114         Around 200  $\mu$ l of haemolymph was then extracted to establish baseline physiological  
115 condition using 1 mL syringes coupled with 25G needles. A sub-sample of 100  $\mu$ l of  
116 haemolymph was then de-proteinased in 100  $\mu$ l of 6 M perchloric acid (PCA), mixed and  
117 stored on dry ice. Muscle samples were excised and immediately frozen in liquid nitrogen.

118         Thereafter, another 150 animals were randomly selected from the discard category  
119 and damage and vigour indexes scored as above as they were placed into indexed tube-sets  
120 (Supplementary Fig. 1) for the short-term survival trial of captive specimens. Sex and length  
121 were not recorded at this stage so as not to delay returning these animals to the sea. Air  
122 temperatures during catch sorting and sampling procedures were recorded, as well as the  
123 overall time taken for the sampling and stocking of the tube-sets to be completed  
124 (Supplementary Tables 1 & 3). Tube-sets containing 150 discarded animals per tow were  
125 enclosed in perforated plastic boxes and returned to the sea suspended on a rope mooring at a  
126 depth (usually ~ 30 m) below the upper lower salinity layer.

127 The *Nephrops* stocked in the suspended tube-sets were left to recover for 48 hours and  
128 after this time the boxes were collected. Sex, CL, damage index and vigour index were then  
129 recorded for all these animals. A random sub-sample of 30 animals per tow were then  
130 sampled for physiology as described above.

### 131 *2.3 Measurement of physiological parameters - Haemolymph and muscle L-lactate*

132 L-lactate concentration was measured following Bergmeyer & Bernt (1974) and  
133 modified by Hill et al. (1991). Samples were thawed, de-proteinized and homogenized using  
134 an Ultra Turrax T25 homogeniser. Homogenates (50 µl) were added to microcentrifuge tubes  
135 (Eppendorf) containing 50 µl of NAD<sup>+</sup> (50 mM), 0.85 ml of hydrazine buffer (0.6 M  
136 hydrazine hydrate, 5.6 mM EDTA, 1 M glycine) at pH 9.5 and 1 unit of lactate  
137 dehydrogenase (LDH, Sigma) and incubated for 2 h at 37 °C. Absorbance readings at 340 nm  
138 (Shimadzu, UV Mini 1240) were converted to concentrations of L-lactate using a calibration  
139 curve constructed from lactic acid standards (0.5-10.0 mM).

### 140 *2.4 Measurement of physiological parameters - ATP and its breakdown products*

141 Nucleotide analysis was performed using High Performance Liquid Chromatography  
142 (HPLC) as described by Ryder (1985) and modified by Albalat et al. (2009). The system  
143 consisted of a Spectra system P4000 pump coupled to a SN4000 autosampler and a UV1000  
144 detector set at 254 nm. A Kinetex 5u C18 100A column 250 x 4.6 mm with an internal  
145 particle of 5 µm was used to conduct the separations (Phenomenex, Cheshire, UK). Standard  
146 curves were prepared from adenosine 5'-triphosphate (ATP), adenosine 5'-diphosphate  
147 (ADP), adenosine 5'-monophosphate (AMP), inosine 5'-monophosphate (IMP), inosine  
148 (HxR) and hypoxanthine (Hx) using concentrations ranging from 0.1 to 1.0 mM (Sigma-  
149 Aldrich, Dorset, UK). The Adenylate Energy Charge (AEC) was calculated based on the  
150 nucleotide profiling of the muscle samples following Atkinson (1968).

151

152

$$AEC = \frac{[ATP] + 0.5 * [ADP]}{[ATP] + [ADP] + [AMP]}$$

153 *2.5 Statistical analyses of physiological data*

154 Differences of L-lactate (in haemolymph and muscle) and AEC ratios between trials  
155 (February, March and June) at time zero and time 48 h were analysed using analysis of  
156 variance (ANOVA) following homogeneity of variance checks using Levene's test. Post-hoc  
157 multiple comparisons (Tukey post-hoc or Games-Howell test according to outcome Levene's  
158 test) were applied to determine statistical differences between samples. Differences between  
159 time zero and time 48 h within each trial were analysed by independent sample t-tests. In all  
160 cases p-values lower than 0.05 were considered statistically significant.

161 *2.6 Predator avoidance behavioural trials using video surveillance*

162 Six, 50 m depth submersible 12 V cameras (VN37CSHR-W36, Visionhitech Co., Ltd,  
163 Korea) were mounted in a custom-built lander frame in a grid layout so that they would cover  
164 a larger area of seabed. The cameras were connected to a multi-channel Digital Video  
165 Recorder (Hybrid 16 DVR, Huviron Ltd., Korea) on-board the research vessel 'Calanus'  
166 which was anchored next to the rig. The rig was deployed on 29<sup>th</sup> July 2015 at 55.805°N  
167 004.882°W at a depth of just under 30 m. Temperature and salinity profiles were recorded  
168 using a Valeport Castaway CTD (Supplementary Fig. 2). The bottom type was sandy-mud  
169 with some shell debris. Discard category *Nephrops* were sorted on board the commercial  
170 fishing vessel following the normal commercial procedures described above and were  
171 transferred to the research vessel as soon as possible once catch sorting was completed. Ten  
172 animals were selected at random and placed in a perforated plastic container and taken  
173 directly to the camera rig by diver. Five deployments were made with animals having been  
174 exposed to varying total air-times and recovery intervals (Supplementary Table 4). The

175 recorded videos were subsequently transferred from the DVR and individual video streams  
176 merged into a composite montage using a custom script written in Matlab. Behaviour of  
177 *Nephrops* and interactions with other organisms were then summarised based on visual  
178 analysis.

### 179 *2.7 Modelling overall survival of discarded Nephrops*

180 Survival proportions of all discarded *Nephrops* were analysed using binomial logistic  
181 regression, with resulting coefficients converted to 95% confidence intervals. Three analyses  
182 were performed: (1) modelling actual survival at 48 h with trawl date, CL, sex, vigour,  
183 damage, infection status, moulting status and females showing mature gonads as explanatory  
184 variables; (2) modelling actual survival at 48 h using only date, sex and carapace length as  
185 explanatory variables (in order to create a model that can be extrapolated to other catch  
186 compositions in each season) and (3) modelling probable survival using the same explanatory  
187 variables as in model 2, where it was assumed that all infected or damaged (class 1 or 2)  
188 animals after 48 h recovery would die shortly after (in order to produce a more conservative  
189 survival estimate as studies such as Wileman *et al.*, 1999 have suggested mortality in  
190 recovering trawled *Nephrops* can take longer than 48 h to stabilise. For each of these models  
191 stepwise regression was performed with non-significant explanatory variables being  
192 removed. All statistical analyses were conducted using R version 3.2.3.

## 193 **3. Results**

### 194 *3.1 General information and fractions of retained and discarded Nephrops*

195 Air temperatures during the trials in February and March were between 7 and 8.5°C  
196 whilst in June temperatures ranged from 12.5°C in the early morning to 15°C by the  
197 afternoon. The time taken to sort the catches on board the fishing vessel was  $83 \pm 9$  min.

198 Sorting times in February and June were slightly longer (between 95-103 min) compared to  
199 March where the sorting time was around 60 min due to the smaller catches. The visual  
200 impression was that total catch volumes (*Nephrops* plus by-catch) were largest in February.  
201 The total numbers of *Nephrops* caught were noticeably lower during the March trial (234-739  
202 tow<sup>-1</sup>) compared to February (1180-1830 tow<sup>-1</sup>) and June (763-1478 tow<sup>-1</sup>). This was due to  
203 the stronger tides during the March trial, conditions that are often associated with reduced  
204 catches of *Nephrops* in this fishery (skipper, personal communication).

205 The overall *Nephrops* discarding rate when expressed as a percentage of the total  
206 number caught was  $29.9 \pm 8.1\%$ . The most abundant size category caught was ‘extra-small’  
207 ( $32.5 \pm 0.9$  mm) followed by ‘small’ ( $37.7 \pm 1.1$  mm), ‘medium’ ( $44.4 \pm 1.4$  mm) and ‘large’  
208 ( $49.3 \pm 1.1$  mm) categories. Animals in the ‘extra-small’ category are marketable and  
209 normally landed by this fishery. The average size of the ‘discard’ category animals was  
210  $28.2 \pm 1.0$  mm which was still above the MCRS and *Nephrops* below the legal size limit only  
211 comprised <1% of the animals being discarded. Significant seasonal-related differences in  
212 discards were also observed with larger animals being rejected in June, compared to March  
213 and February (Fig. 1A). This pattern is explicable because the percentage of females in the  
214 discard fraction was higher in June and the majority of these animals were soft and had  
215 mature gonads (Fig. 2B). These animals have low market value and are therefore normally  
216 discarded (skipper, personal communication).

### 217 *3.2 Physiological condition and short-term survival of captive discarded Nephrops*

218 Overall, discard fraction *Nephrops* were predominantly in vigour category ‘C’ (51%)  
219 immediately after trawling (Fig. 1A) while animals in top vigour categories A+B accounted  
220 for around 25%. After a 48 h recovery period, the proportion of discarded animals in the top  
221 vigour categories (A+B) had increased to 64% while 12% were moribund or dead. From a  
222 seasonal perspective, a higher proportion of the *Nephrops* were classified as being in vigour

223 category A after the 48 h recovery period in February (Fig. 1B). Moreover, the number of  
224 moribund animals was lowest although the numbers of prawns dying was actually higher. In  
225 June more animals were in vigour categories B+C after 48 h recovery but the number which  
226 had died was lower than in other months.

227         Immediately after trawling 87% of the discard fraction *Nephrops* had no visible signs  
228 of damage. There was also no obvious seasonal effect on the damage scores with values  
229 being low throughout (data not shown) neither did damage appear to be related to sex, size or  
230 moulting stage of the animals. After the 48 h recovery period there was an increase in the  
231 proportion of animals with some visible damage (13% at T0 to 18% at T48). Potential  
232 contributing factors could include handling the animals in and out of the tube-sets or moving  
233 the boxes in and out of the water. Although in this study, damage percentages were low,  
234 damage did have an impact on the recovery potential of the animals. As shown in Fig. 3A, the  
235 proportion of dead and moribund animals after 48 h recovery increased with damage. Indeed,  
236 more than 50% of the *Nephrops* which were classified as being in damage category 2  
237 immediately after trawling were dead or moribund after 48 h.

238         The percentages of prawns with *Hematodinium* infection varied seasonally; 10.2% of  
239 animals infected in February, 12.9% in March but only 0.4% in June. Infected animals did  
240 not show as good recovery potential as un-infected animals. After 48 h recovery more than  
241 50% of animals classified as visually infected immediately after trawling were dead or  
242 moribund (Fig. 3B).

243         Baseline physiological data for prawns immediately post-catch were indicative of  
244 capture stress with elevated muscle <sub>L</sub>-lactate (Fig. 3A) and low AEC ratios indicative of  
245 anaerobic respiration (**Error! Reference source not found.**B). In June, muscle <sub>L</sub>-lactate  
246 values after 48 h were significantly higher than in February or March (Fig. 3A) but seasonal  
247 effects were not so obvious in post-recovery AEC ratios (Fig. 3B). Despite some indications

248 that physiological recovery was poorer in June (lower percentage of animals in vigour  
249 category A and significantly higher muscle L-lactate concentration) the percentage of dead  
250 animals after 48 h was actually lowest in that month (Fig. 2B).

### 251 3.3 Predator avoidance behavioural observations on released discard fraction *Nephrops*

252 During the behavioural trial the air temperature rose from 9.5°C in the early morning  
253 to around 17°C by 10:00 whilst CTD casts showed a layer of lower salinity water reaching  
254 down to around 5 m depth (Fig. S2). All animals used were in the discard size class (22–  
255 32 mm CL) and all test animals, except one, were in damage category 0. Prawns caught in the  
256 morning were all in vigour index A, B or C after transfer to the research vessel. The vigour  
257 indices of *Nephrops* caught in the afternoon were slightly lower (mainly B/C) (Table S4). In  
258 the main deployments (2 to 5), delays in returning the animals to the sea were relatively short  
259 (30–45 mins to re-immersion). The exposure to seawater during their descent seemed to act as  
260 a stimulant and nearly all the *Nephrops* (38/40) exhibited almost immediate righting action  
261 and began moving around when placed on the seabed (Table S5). The majority of these  
262 animals appeared to exhibit normal locomotory behaviour and dispersed out of the camera  
263 view within a few minutes. In the initial test deployment the animals had been held on the  
264 research vessel in air for around 90 mins while the equipment and dive team were being  
265 prepared. These *Nephrops* were mostly in vigour category C or were moribund when taken  
266 down to the camera rig. Some of these animals took up to 10 minutes to recover and this time  
267 period was sufficient to attract several shore crabs (*Carcinus maenas*) and squat lobsters  
268 (likely *Munida rugosa*) which attacked any un-righted *Nephrops*. Although several prawns  
269 exhibited tail-flip escape responses this tended to be a single or double flip rather than a  
270 series which would have carried them away from the predators. At least one prawn would

271 probably have failed to survive (the individual in question being attacked by two crabs and  
272 eventually dragged out of the observation arena (Table S5).

### 273 3.4 Modelling overall survival of discarded *Nephrops*

274 Of the available co-variables (date, CL, sex, damage, vigour, softness, infection and  
275 gonad status) only date, damage index, infection status and moulting stage were found to be  
276 statistically significant explanatory variables for 48 h survival overall. Table 1 shows  
277 confidence intervals for the baseline probability of survival in each season and also the  
278 increase in the odds ratio of death according to each contributory factor. Only date was found  
279 to be statistically significant for the model of likely long-term survival (in this more  
280 conservative model it was assumed that damaged, infected or soft animals would all die in the  
281 longer-term, Table 2).

282 As a caveat it should also be noted that there was an increase in the proportion of  
283 animals with some damage during the 48 h recovery period from 12.8% at T0 to 18.0% at  
284 T48. It was not possible to correct survival estimates for this because the additional damage  
285 could have occurred at any time during the 48 h recovery, or during inspection of the animals  
286 after this time. Nevertheless, such accidentally induced additional damage would only  
287 decrease survival meaning that the estimated survival rates will be conservative.

## 288 4. Discussion

289 Although other studies have described the annual emergence of recently moulted  
290 female *Nephrops* in spring and early summer (Milligan et al., 2009) this is the first time that  
291 discarding of soft females as a low quality product has been reported. Similarly, although  
292 several studies have reported the seasonal patterns of *Nephrops* infection by *Hematodinium*  
293 (Field et al., 1992; Stentiford et al., 2001; Beevers et al., 2012), this is the first time that it has  
294 been observed that fishers are able to recognise symptomatic patent infected animals (on the

295 basis of their body colour) and that these animals are then being discarded. The choice to  
296 discard soft or infected animals is driven by the fact that this particular fishery is serving the  
297 live market and prioritising product quality over volume. Although such high-grading is often  
298 considered to be an undesirable fisheries practice, the actual impacts will vary depending on  
299 the survivability of the discarded fraction.

300 In the present study more than 90% of the discarded *Nephrops* were alive after 48 h  
301 re-immersion in tube-sets. These values are higher than reported in most other post-discard  
302 studies in *Nephrops* trawl fisheries (e.g. Castro et al., 2003; Méhault et al., 2016). Guéguen  
303 and Charuau (1975) reported a survival rate of ~25% after three days recovery in cages  
304 immersed at around 60 m depth, following aerial exposure of 90 min but later studies have  
305 reported higher survival rates ranging from 31% (Charuau et al., 1982); 30-48% (Castro et  
306 al., 2003) to 45-65% (Méhault et al., 2016). There are multiple potential reasons for the wide  
307 disparities in estimates of post-discard survival including differences in the fishing gears and  
308 practices; differences in water and air temperatures and the length of air exposure and  
309 differences in how the catches are handled. The higher survival rates in the present study are  
310 likely related to the specific nature of this Clyde fishery that prioritises quality over volume  
311 resulting in short tows and rapid catch sorting.

312 Nevertheless the lack of standardisation in experimental designs does cause problems  
313 when comparing results between studies (Méhault et al., 2016, ICES, 2014). This is apparent  
314 not only in terms of the number of vessels which should be used (most survival trials  
315 conducted to date have been carried out in single vessels felt to be ‘representative’ of the  
316 wider fleet practices) but also in relation to the length of time that experiments need to be run  
317 for (Wassenberg and Hill, 1993; Castro *et al.*, 2003; Campos *et al.*, 2015). A report by  
318 STECF (2015) suggested observations on *Nephrops* recovery should be continued until  
319 mortality rates have stabilised which in some studies took as long as 10-14 days (Wileman et

320 al., 1999). Measurements in the present study and in Albalat *et al.* (2010) suggest that  
321 physiological recovery to resting values occurs within 24-48 h. Delayed mortality thus  
322 appears to be mainly the result of physical damage (Wileman *et al.* 1999). Our model 2  
323 survival estimate therefore assumed that any animals which were damaged or infected after  
324 48 h recovery would die over the following few days and so this represents a conservative  
325 assumption of survival of around 63% (the lowest 95% CI). This approach might prove  
326 useful as establishing mortality curves over 10-14 days means that recovery cages must be  
327 inspected regularly. Regular inspection requires either raising them to the surface, which  
328 causes increased aerial exposure and stress, or the use of divers (feasible down to about 30 m)  
329 or potentially inspection using remotely-operated vehicles (ROV). Monitoring of recovery  
330 cages by divers has been used in some previous studies (Wileman *et al.*, 1999) but  
331 considerably increases experimental costs. A final alternative is to allow animals to recover in  
332 tanks on-board fishing vessels or ashore (Edwards and Bennett, 1980; ICES, 2014). Whether  
333 tanks, however, adequately replicate natural recovery conditions is debateable especially  
334 where it is difficult to control factors such as temperature and salinity. Exposure to low  
335 salinity water in particular is known to cause stress to *Nephrops* (Harris and Ulmestrand,  
336 2004). Because of these problems most studies on *Nephrops* have not monitored individual  
337 survival over time but have used start-end comparisons in recovery cages deployed for  
338 periods of between 2-9 days (Castro *et al.*, 2013; Campos *et al.*, 2015).

339         Because mortality in discarded *Nephrops* seems to be more related to physical  
340 damage than physiological stress, changes in fishing practice aimed at reducing damage  
341 should increase post-catch survival. Implementing such changes may be difficult within a  
342 commercial fishery context. For example the use of short tows, which appears beneficial in  
343 this fishery, may not be acceptable in other *métiers* where catch volume is more important  
344 and tow durations are up to four hours (Milligan *et al.*, 2009; Fox, 2010). Improved survival

345 of discarded *Nephrops* has also been reported with some gear modifications, such as in-trawl  
346 sorting grids and SELTRA panels (STECF, 2015). This may be related to reductions in the  
347 overall bulk of catch in the cod-end leading to reduced damage (Broadhurst et al., 2009).  
348 Nevertheless, these results remain to be confirmed as the higher survivals reported could also  
349 have a seasonal, rather than gear modification, basis (STECF, 2015).

350         Although the majority of discard survivability trials have allowed animals to recover  
351 in predator-free enclosures (Wileman et al., 1999; Bergmann and Moore, 2001; Castro et al.,  
352 2003; Revill et al., 2005; Mandelman and Farrington, 2007; Rulifson, 2007; Enever et al.,  
353 2009; Campos et al., 2015), it is widely recognised that this may over-estimate post-discard  
354 survival (Raby et al., 2014). Animals returned to the sea must escape potential predators and,  
355 in the case of *Nephrops*, seek protection in existing, or newly excavated burrows. In the  
356 present study around 80% of the discarded *Nephrops* were in vigour categories A-C  
357 immediately after being trawled (Fig. 3A) and results from our underwater observations  
358 suggested that these animals could right themselves, begin active movement and adopt  
359 natural defensive postures and escape responses within a few minutes of being returned to the  
360 seabed. On the other hand, animals in a moribund condition, as a result of longer air  
361 exposure, took up to 10 minutes to recover on the seabed. This time period was sufficient to  
362 attract predators, such as shore crabs and squat lobsters, which were observed to successfully  
363 capture at least one of these *Nephrops*. The metabolic capacity of moribund *Nephrops* thus  
364 appears to be compromised to the point where they struggle to make repeated escape tail-flips  
365 when attacked by predators (Gornik et al., 2008). Since moribund animals made up around  
366 20% of the discard category immediately after trawling (Fig. 3A) these animals are likely to  
367 be more vulnerable to predation after reaching the seabed. Unfortunately our behavioural  
368 observations only extended over the immediate post-release period because we did not want  
369 to use fences around the camera arena which would exclude benthic predators. The relatively

370 rapid attraction of predators to the discarded *Nephrops*, as also observed by Bergmann *et al.*  
371 (2002), suggests that future studies need to place more emphasis on such behavioural  
372 interactions. Discarded *Nephrops* might also be vulnerable to predation during the descent to  
373 the seabed but based on measurements of sinking rates and estimates of midwater predation,  
374 Bergmann *et al.* (2002) suggested that most discards in the Firth of Clyde would reach the  
375 seabed. For longer term survival studies STECF (2015) have suggested that tag and release  
376 programs may be required. Such programs would have to be run over extensive periods, tag  
377 loss rates evaluated, large numbers of *Nephrops* marked and a large percentage of the fishing  
378 vessels in the area equipped to detect the presence of the internal tags which would have to be  
379 used. The costs involved in such large-scale discard-survival programs are such that they may  
380 only be worth undertaking for animals with higher conservation concern and which are  
381 occasionally caught in *Nephrops* trawls, such as the flapper skate *Dipturus intermedia* (Neat  
382 *et al.*, 2015).

383         Given the many factors which vary between fishing fleets, grounds, seasons and even  
384 individual vessels, producing realistic survival estimates for animals discarded in commercial  
385 fisheries remains extremely challenging. This study was conducted in a commercial vessel  
386 thought to be representative of practices in the Firth of Clyde trawl fishery that targets the  
387 live market. Further studies may be required to confirm that the data presented are  
388 representative of the wider fleet.

389         The fishery studied here is characterised by day vessels that perform short tows  
390 (typically < 2.5h) and practices which prioritise the quality of the animals over catch volume  
391 (Combes, 2007). This makes this fishery quite specialised compared with the majority of  
392 *Nephrops* trawling where tow durations and sorting times tend to be longer and the catches  
393 larger. Data from this study should therefore not be extrapolated to other trawl-based

394 *Nephrops* fisheries where further studies will be required to estimate the survival of discarded  
395 animals.

## 396 **Acknowledgments**

397 This research was funded by a grant from Fisheries Innovation Scotland (FIS007, 2015). The  
398 authors would like to acknowledge the enthusiastic co-operation of Ian Wightman, the  
399 skipper of the Eilidh Anne; the skipper and crew of the SAMS research vessel ‘Calanus’ and  
400 the NERC Scientific Diving team (based at SAMS) without whom this work would not have  
401 been possible.

## 402 **References**

- 403 Albalat, A., Gornik, S.G., Atkinson, R.J.A., Coombs, G.H., Neil, D.M. 2009. Effect of  
404 capture method on the physiology and nucleotide breakdown products in the Norway  
405 lobster (*Nephrops norvegicus*). Mar. Biol. Res. 5, 441-450.
- 406 Albalat, A., Sinclair, S., Laurie, J., Taylor, A., Neil, D. 2010. Targeting the live market:  
407 Recovery of Norway lobsters *Nephrops norvegicus* (L.) from trawl-capture as assessed  
408 by stress-related parameters and nucleotide breakdown. J. Exp. Mar. Biol. Ecol. 395,  
409 206-214.
- 410 Atkinson, D.E. 1968. Energy charge of the adenylate pool as a regulatory parameter.  
411 Interaction with feedback modifiers. Biochem. 7, 4030-4034.
- 412 Beevers, N.D., Kilbride, E., Atkinson, R.J.A., Neil, D.M. 2012. *Hematodinium* infection  
413 seasonality in the Firth of Clyde (Scotland) *Nephrops norvegicus* population: a re-  
414 evaluation. Dis. Aqua. Org. 100, 95-104.
- 415 Bergmann, M., Moore, P.G. 2001. Survival of decapod crustaceans discarded in the  
416 *Nephrops* fishery of the Clyde Sea area, Scotland. ICES J. Mar. Sci. 58, 163-171.

- 417 Bergmann, M., Wieczorek, S.K., Moore, P.G., Atkinson, R.J.A. 2002. Utilisation of  
418 invertebrates discarded from the *Nephrops* fishery by variously selective benthic  
419 scavengers in the west of Scotland. Mar. Ecol. Prog. Ser. 233, 185-198.
- 420 Bergmeyer, H.U., Bernt, E. 1974. Determination of glucose with glucose oxidase and  
421 peroxidase, in: Bergmeyer, H.U. (Ed.), Methods of enzymatic analysis. Verlag Chemie-  
422 Academic Press, Weinheim/Bergstr. (Germany), pp. 1205-1215.
- 423 Borges, L. 2015. The evolution of a discard policy in Europe. Fish Fish. 16, 534-540.
- 424 Broadhurst, M.K., Millar, R.B., Uhlmann, A.S. 2009. Using a double codend to reduce  
425 discard mortality. ICES J. Mar. Sci. 66, 2077-2081.
- 426 Campos, A., Fonseca, P., Pilar-Fonseca, T., Leocádio, A.M., Castro, M. 2015. Survival of  
427 trawl-caught Norway lobster (*Nephrops norvegicus* L.) after capture and release —  
428 Potential effect of codend mesh type on survival. Fish. Res. 172, 415-422.
- 429 Castro, M., Araújo, A., Monteiro, P., Madeira, A.M., Silvert, W. 2003. The efficacy of  
430 releasing caught *Nephrops* as a management measure. Fish. Res. 65, 475-484.
- 431 Catchpole, T.L., Frid, C.L.J., Gray, T.S. 2005a. Discarding in the English north-east coast  
432 *Nephrops norvegicus* fishery: the role of social and environmental factors. Fish. Res.  
433 72, 45-54.
- 434 Catchpole, T.L., Frid, C.L.J., Gray, T.S. 2005b. Discards in North Sea fisheries: causes,  
435 consequences and solutions. Mar. Policy 29, 421-430.
- 436 Charuau, A., Morizur, Y., Rivoalen, J.J. 1982. Survie des rejets de *Nephrops norvegicus* dans  
437 le Golfe de Gascogne et en mer Celtique. ICES C.M. 1982/B:13. 6 pp.
- 438 Condie, H.M., Grant, A., Catchpole, T.L. 2014. Incentivising selective fishing under a policy  
439 to ban discards; lessons from European and global fisheries. Mar. Policy 45, 287-292.
- 440 Combes, J. 2007. Clyde Environment and Fisheries Review and Sustainable Supply Chain  
441 Project Report. Clyde Fisheries Development Project. 196 pp.

- 442 Edwards, E.S., Bennett, D.B. 1980. Survival of discarded *Nephrops*. ICES CM 1980/K:10.  
443 6 pp.
- 444 Eliassen, S.Q., Papadopoulou, K.-N., Vassilopoulou, V., Catchpole, T.L. 2014. Socio-  
445 economic and institutional incentives influencing fishers' behaviour in relation to  
446 fishing practices and discard. ICES J. Mar. Sci. 71, 1298-1307.
- 447 Enever, R., Catchpole, T.L., Ellis, J.R., Grant, A. 2009. The survival of skates (Rajidae)  
448 caught by demersal trawlers fishing in UK waters. Fish. Res. 97, 72-76.
- 449 Field, R.H., Chapman, C.J., Taylor, A.C., Neil, D.M., Vickerman, K. 1992. Infection of the  
450 Norway lobster *Nephrops norvegicus* by a *Hematodinium*-like species of dinoflagellate  
451 on the west coast of Scotland. Dis. Aquat. Org. 13, 1-15.
- 452 Fox, C.J. 2010. West coast fishery trials of a twin rig *Nephrops* trawl incorporating a large  
453 mesh top sheet for reducing commercial species gadoid bycatch. Scottish Industry  
454 Science Partnership Report 03. Marine Scotland Science, 39 pp.
- 455 Fox, C.J. 2014. A workshop to address the issues surrounding a discarding ban in the Scottish  
456 *Nephrops* fisheries. Marine Alliance for Science and Technology for Scotland, 37 pp.  
457 doi:10.13140/2.1.2661.2802
- 458 Fox, C.J., Valcic, L., Veszalovski, A. (2015) A pilot study to define the footprint and  
459 activities of Scottish inshore fisheries by identifying target fisheries, habitats and  
460 associated fish stocks. Evidence gathering in support of sustainable Scottish inshore  
461 fisheries: Work Package 4 Final Report SRS004SIF, SRSL, Oban, 193 pp.
- 462 Gaten, E., Moss, S., and Johnson, M.L. 2013. The reniform reflecting superposition  
463 compound eyes of *Nephrops norvegicus*: Optics, susceptibility to light-induced  
464 damage, electrophysiology and a ray tracing model, in: Magnus, L.J., Mark, P.J. (Eds.),  
465 Advances in Marine Biology, Volume 64: Academic Press, pp. 107-148.

466 Gornik, S.G., Albalat, A., Atkinson, R.J.A., Coombs, G.H., and Neil, D.M. 2008. The time  
467 course of early post-mortem biochemical processes in the abdominal muscle of a  
468 commercially important decapod crustacean (*Nephrops norvegicus*): implications for  
469 post-catch processing. Mar. Freshwater Behav. Physiol. 41, 241-256.

470 Guéguen, J., Charuau, A. 1975. Essai de détermination du taux de survie des langoustines  
471 hors taille rejetées lors des opérations de pêche commerciale. ICES CM 1975/K:12.

472 Hamilton, K.M., Shaw, P.W., and Morritt, D. 2009. Prevalence and seasonality of  
473 *Hematodinium* (Alveolata: Syndinea) in a Scottish crustacean community. ICES J. Mar.  
474 Sci. 66, 1837-1845.

475 Harris, R.R., Andrews, M.B. 2005. Physiological changes in the Norway lobster *Nephrops*  
476 *norvegicus* (L.) escaping and discarded from commercial trawls on the West Coast of  
477 Scotland: II. Disturbances in haemolymph respiratory gases, tissue metabolites and  
478 swimming performance after capture and during recovery. J. Exp. Mar. Biol. Ecol. 320,  
479 195-210.

480 Harris, R.R., and Ulmestrand, M. 2004. Discarding Norway lobster (*Nephrops norvegicus* L.)  
481 through low salinity layers – mortality and damage seen in simulation experiments.  
482 ICES J. Mar. Sci. 61, 127-139.

483 Hill, A.D., Taylor, A.C., Strang, R.H.C. 1991. Physiological and metabolic responses of the  
484 shore crab *Carcinus maenas* (L.) during environmental anoxia and subsequent  
485 recovery. J. Exp. Mar. Biol. Ecol. 150, 31-50.

486 ICES 2013. Report of the Working Group on the Assessment of Demersal Stocks in the  
487 North Sea and Skagerrak (WGNSSK), 24-30 April 2013, ICES Headquarters,  
488 Copenhagen, CM 2013/ACOM:13. 1435 pp.

489 ICES 2014. Report of the ICES Workshop on Methods for Estimating Discard Survival  
490 (WKMEDS), 17-21 February 2014, ICES Headquarters, Copenhagen, CM  
491 2014/ACOM:51. 117 pp.

492 Kaiser, M.J., Spencer, B.E. 1994. Fish scavenging behaviour in recently trawled areas. Mar.  
493 Ecol. Progress Ser. 112, 41-49.

494 Lund, H.S., Wang, T., Chang, E.S., Pedersen, L.F., Taylor, E.W., Pedersen, P.B., McKenzie,  
495 D.J. 2009. Recovery by the Norway lobster *Nephrops norvegicus* (L.) from the  
496 physiological stresses of trawling: Influence of season and live-storage position. J. Exp.  
497 Mar. Biol. Ecol. 373, 124-132.

498 Mandelman, J.W., Farrington, M.A. 2007. The estimated short-term discard mortality of a  
499 trawled elasmobranch, the spiny dogfish (*Squalus acanthias*). Fish. Res. 83, 238-245.

500 Marine Scotland, 2015. Implementation of the demersal landing obligation in 2016. Marine  
501 Scotland guidance for Scottish fishing vessels. Report produced by The Scottish  
502 Government accessible at <http://www.gov.scot/Resource/0049/00490271.pdf>

503 Méhault S., Morandeau F., Kopp D., 2016. Survival of discarded *Nephrops norvegicus* after  
504 trawling in the Bay of Biscay. Fish. Res. 83, 396–400.

505

506 Milligan, R.J., Albalat, A., Atkinson, R.J.A., Neil, D.M. 2009. The effects of trawling on the  
507 physical condition of the Norway lobster *Nephrops norvegicus* in relation to seasonal  
508 cycles in the Clyde Sea area. ICES J. Mar. Sci. 66, 488-494.

509 Neat, F., Pinto, C., Burrett, I., Cowie, L., Travis, J., Thornburn, J., Gibb, F.M., Wright, P.J.  
510 2015. Site fidelity, survival and conservation options for the threatened flapper skate  
511 (*Dipturus cf intermedia*). Aquat. Conserv. 25, 6-20.

512 Raby, G.D., Packer, J.R., Danylchuk, A.J., Cooke, S.J. 2014. The understudied and  
513 unappreciated role of predation in the mortality of fish released from fishing gears. *Fish*  
514 *Fish.* 15, 489-505.

515 Revill, A. S., Dulvy, N. K., .Holst, R. 2005. The survival of discarded lesser-spotted dogfish  
516 (*Scyliorhinus canicula*) in the Western English Channel beam trawl fishery. *Fish. Res.*  
517 71, 121-124.

518 Ridgway, I.D., Taylor, A.C., Atkinson, R.J.A., Chang, E.S., Neil, D.M. 2006a. Impact of  
519 capture method and trawl duration on the health status of the Norway lobster, *Nephrops*  
520 *norvegicus*. *J. Exp. Mar. Biol. Ecol.* 339, 135-147.

521 Ridgway, I.D., Taylor, A.C., Atkinson, R.J.A., Stentiford, G.D., Chang, E.S., Chang, S.A.,  
522 Neil, D.M. 2006b. Morbidity and mortality in Norway lobsters, *Nephrops norvegicus*:  
523 physiological, immunological and pathological effects of aerial exposure. *J. Exp. Mar.*  
524 *Biol. Ecol.* 328, 251-264.

525 Rulifson, R.A. 2007. Spiny dogfish mortality induced by gill-net and trawl capture and tag  
526 and release. *North Am. J. Fish. Manag.* 27, 279-285.

527 Ryder, J.M. 1985. Determination of adenosine triphosphate and its breakdown products in  
528 fish muscle by high-performance liquid chromatography. *J. Agricult. Food Chem.* 33,  
529 678-680.

530 Sigurðardóttir, S., Stefánsdóttir, E.K., Condie, H., Margeirsson, S., Catchpole, T.L., Bellido,  
531 J.M., Eliassen, S.Q., Goñi, R., Madsen, N., Palialexis, A., Uhlmann, S.S.,  
532 Vassilopoulou, V., Feekings, J., Rochet, M.-J. 2015. How can discards in European  
533 fisheries be mitigated? Strengths, weaknesses, opportunities and threats of potential  
534 mitigation methods. *Mar. Policy* 51, 366-374.

535 STECF 2015. Landing obligation - Part 5 (demersal species for NWW, SWW and North  
536 Sea), JRC Scientific Technical and Economic Committee for Fisheries, Publications  
537 Office of the European Union, Luxembourg, EUR 27407 EN, JRC 96949. 65 pp.

538 Stentiford, G.D., Neil, D.M., Atkinson, R.J.A. 2001. The relationship of *Hematodinium*  
539 infection prevalence in a Scottish *Nephrops norvegicus* population to season, moulting  
540 and sex. ICES J. Mar. Sci. 58, 814-823.

541 Ungfors, A., Bell, E., Johnson, M.L., Cowing, D., Dobson, N.C., Bublitz, R., Sandell, J.  
542 2013. *Nephrops* fisheries in European waters, in: Magnus, L.J., Mark, P.J. (Eds.),  
543 Advances in Marine Biology, Volume 64, Academic Press, pp. 247-314.

544 Wassenberg, T.J., Hill, B.J. 1993. Selection of the appropriate duration of experiments to  
545 measure the survival of animals discarded from trawlers. Fish. Res. 17, 343-352.

546 Wileman, D.A., Sangster, G.I., Breen, M., Ulmestrand, M., Soldal, A.V., Harris, R.R. 1999.  
547 Roundfish and *Nephrops* survival after escape from commercial gear, Final report, EC  
548 Contract No: FAIR-CT95-0753. 240 pp.

549  
550  
551

**Table 1:** 95% confidence intervals for results of model 1. Results are expressed as baseline probability of mortality in each season, and increase in odds ratio of mortality according to each contributory factor.

<b>Season</b>	<b>95% confidence interval of probability of survival (healthy un-damaged <i>Nephrops</i>)</b>
Feb	93.4 – 97.1%
Mar	95.6 – 98.8%
June	98.8 – 99.9%
<b>Category</b>	<b>Increase in odds ratio of death comparing undamaged and uninfected animals with other categories</b>
Damage class 1	1.3 – 5.4 times
Damage class 2	2.0 – 35.8 times
Visually Infected	2.2 – 11.4 times
Soft	3.1 – 31.3 times

**Table 2:** 95% confidence interval for probability of survival for each season.

<b>Model</b>	<b>February</b>	<b>March</b>	<b>June</b>
(1) Actual survival, based on percentage of animals alive after 48 h recovery	90.1 – 95.8%	91.8 – 96.9 %	96.1 – 98.9 %
(2) Likely survival, assuming damaged or infected animals will not survive in the longer term	68.0 – 77.7 %	63.4 – 73.8 %	81.3 – 88.0 %

## Fig. legends

**Fig. 1.** (A) Size distribution of discarded *Nephrops* in the discard fraction by carapace length (mm) in the different trials (B) Percentage of 'soft' prawns in the discard fraction. From the soft category discarded in June the percentage of females and also the percentage of soft females that showed mature gonads are also shown. Vertical whiskers indicate mean  $\pm$  S.E.

**Fig. 2.** (A) Percentage of discarded *Nephrops* classified according to the vigour index just after trawling (time 0) and 48 hours after recovery at sea from all discarded animals used in the survival trials (B) Percentage of discarded *Nephrops* classified according to the vigour index 48 h after recovery period in each trial.

**Fig. 3.** Percentage of discarded *Nephrops* alive (vigour A-C), moribund or dead after 48 hours recovery categorised by their initial (A) external damage and (B) visual infection with *Hematodinium*.

**Fig. 4.** (A) Abdominal muscle and haemolymph L-lactate concentrations and (B) AEC ratio in the muscle of discarded *Nephrops* immediately after capture and after 48 h recovery. Values represent the mean  $\pm$  S.E. of n=30 (for T0) and n=60 (for T48) animals for each trial. \* indicates significant differences ( $p < 0.05$ ) between T0 and T48 while different letters indicate significant differences between trials. For reference, AEC values in rested *Nephrops* have been reported as being  $>0.8$  (Albalat et al., 2010).

**Fig 1.**

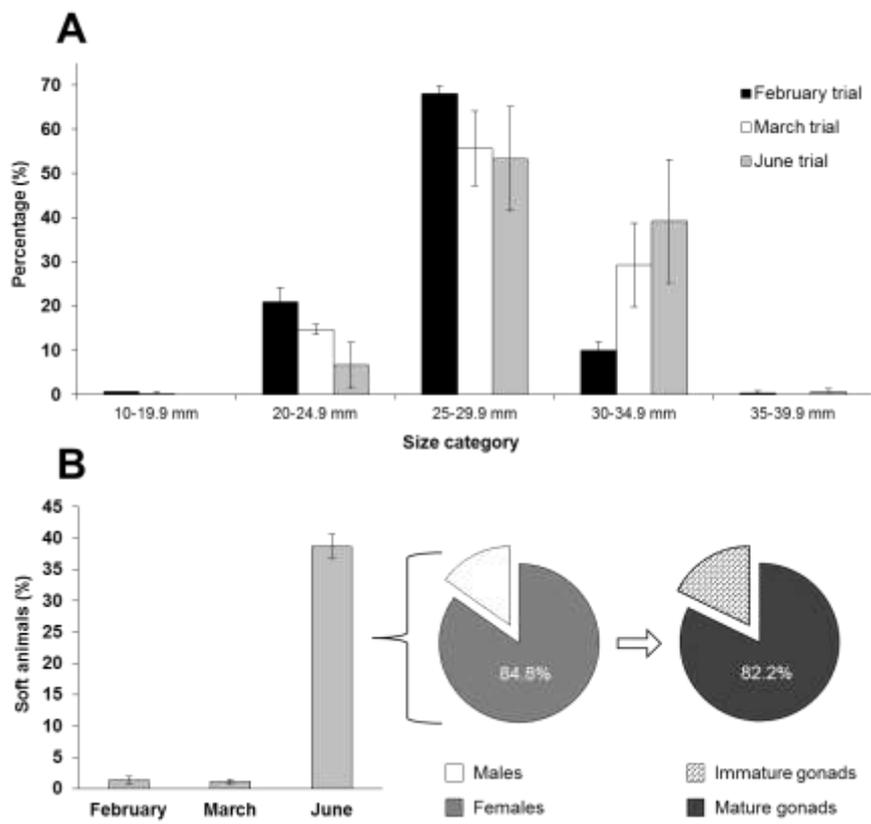
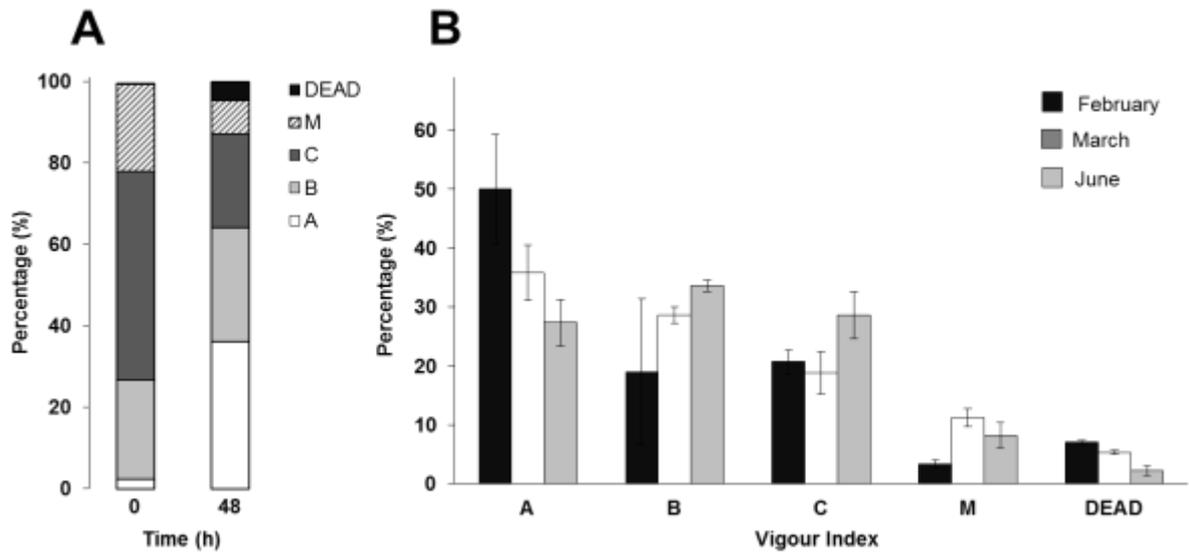


Fig. 2



**Fig. 3**

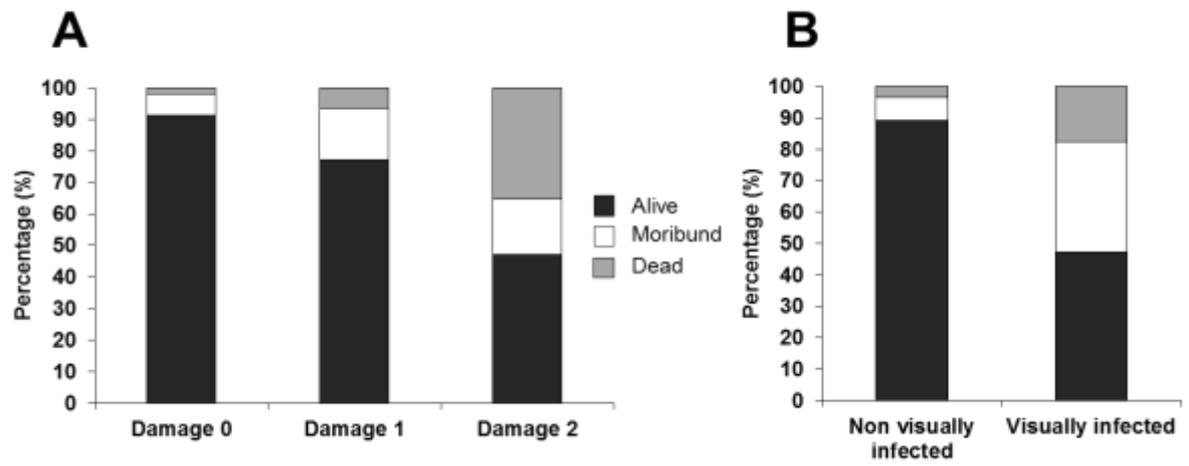
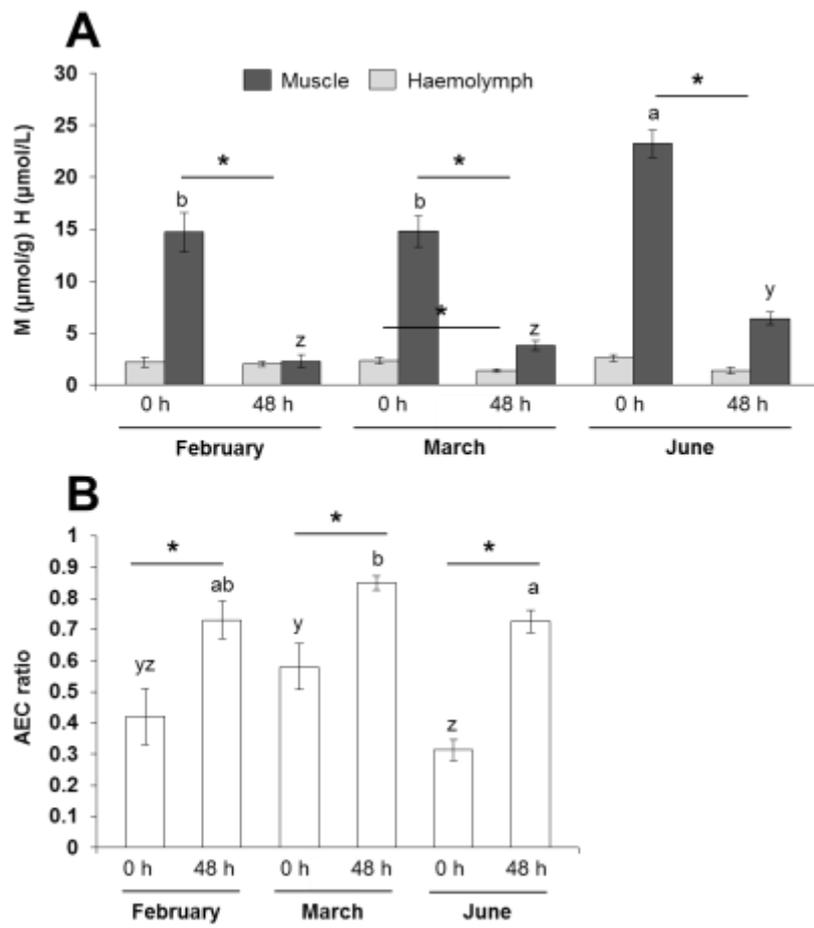


Fig. 4



## Supplementary material

The following supplementary material is available online. **Table S1** General information and environmental conditions during the trawls and initial sampling of *Nephrops* for the recovery trials. **Table S2** General information and environmental conditions during the recovery of tubed animals. **Table S3**. Definition of the different Vigour Index categories used. **Table S4**: Deployment details (times UTC) of the behavioural studies. **Table S5**: Summary of video clips accompanying this report. **Fig. S1**. A representative tube-set used for the survival trial, a tube-set box closed ready to be re-immersed for survival trial. **Fig. S2**. Temperature and salinity water column profiles from the camera deployment locations.

**Table S1.** General information and environmental conditions during the trawls and initial sampling of *Nephrops* for the recovery trials.

Date	11/02/15		24/03/15			16/06/15		
	1	2	1	2	3	1	2	3
Tow								
Weather	Cloudy	Cloudy	Cloudy	Sunny	Sunny	Cloudy	Sunny	Cloudy
Start time trawl	08:00	11:25	07:45	11:00	14:11	07:47	11:35	15:10
End time trawl	09:15	13:00	09:06	12:30	15:58	09:25	13:13	16:24
Trawl duration (min)	75	95	81	90	107	98	98	74
Trawl start	55.757°N 4.895°W	55.800°N 4.884°W	55.799°N 4.882°W	55.802°N 4.888°W	55.820°N 4.899°W	55.814°N 4.903°W	55.803°N 4.893°W	55.802°N 4.719°W
Trawl end	55.128°N 4.889°W	55.127°N 4.888°W	55.129°N 4.889°W	55.129°N 4.890°W	55.156°N 4.900°W	55.036°N 4.070°W	55.031°N 4.901°W	55.005°N 4.908°W
Start trawl depth (m)	35	55	60	80	66	86	85	63
End trawl depth (m)	60	60	80	60	60	84	88	96
Towing Speed (knots)	2.1	2.1	2.1	2.1	2.2	2.0	2.1	2.0
Air temp (°C)	7.1	7.4	7.6	8.5	8.4	12.5	14.0	15.0
Water Surface Temp (°C)	7.9	7.6	7.5	7.9	7.9	11.8	12.0	12.0
Water Bottom Temp (°C)	8.0	8.0	7.3	7.3	7.3	-	-	-
Salinity water Surface (PSS)	32.6	32.2	32.0	31.7	31.7	-	-	-
Salinity water Bottom (PSS)	32.8	33.2	33.1	32.9	32.9	-	-	-
Time animals arrived on-board	09:20	13:08	09:18	12:41	16:11	09:38	13:27	16:35
Time animals put back in water	11:22	14:30	10:32	13:36	17:00	11:13	14:44	18:27
Sorting time (min)	122	84	74	55	49	95	77	112
Depth animals were returned to sea (m)	25	25	30	28	25	36	35	35
Location animals returned to sea to recover		55.132°N 4.874°W		55.778°N 4.895°W			55.777°N 4.894°W	

**Table S2.** Characteristics of the different Vigour Index categories used in the trials.

Vigour A	Vigour B	Vigour C	Moribund	Dead
Animal displays defensive posture, and/or vigorous and possibly prolonged (>10 flips) tail flipping	Bouts of tail flipping are less frequent and vigorous, lasting for shorter periods (<10 flips approx). Animal less likely to adopt defensive posture	Bouts of tail flipping are infrequent and do not normally exceed 5 flips per bout (approx)	Tail flips are very infrequent and weak and are limited to 1 or 2 flips per bout	No limb movements even if animal is stimulated
Antennae and claws are held high and may be ‘waved’ by the animal	Antennae and claws are raised but moved less vigorously	Antennae and claws are drooped but animal can raise them for short periods (a few seconds)	Antennae and claws are drooped and animal is unable to raise them	
Tail is held rigid upright in an angle (when animal is not tail flipping). Related to reflex response.	Tail is held horizontally retains some tension. Related to reflex response.	Tail is drooped and retains very little tension. Related to reflex response.	Tail is limp. Related to reflex response.	
Walking legs are strong and animal will often right itself	Walking legs are moved but animal normally cannot right itself	Walking legs are moved by the animal but it will not be able to right itself	Walking legs are moved very slowly if animal is stimulated	

**Table S3.** General information and environmental conditions during the recovery of tubed animals.

Month	13/02/15		26/03/15			18/06/15		
	1	2	1	2	3	1	2	3
Air temp (°C)	6.5	7.5	6.4	8.3	9.7	12.0	11.5	11.0
Water temp Surface (°C)	7.5	7.5	7.5	7.5	7.5	11.5	11.0	11.0
Water temp Bottom (°C)	8.1	8.1	7.4	7.4	7.4	-	-	-
Salinity water Surface (PSS)	32.6	32.6	31.3	31.3	31.3	-	-	-
Salinity water Bottom (PSS)	33.2	33.2	32.9	32.9	32.9	-	-	-
Time animals arrive on-board	10:09	10:07	09:45	13:04	13:04	09:51	11:30	14:25
Time start sampling	10:15	11:42	09:50	13:23	14:04	09:53	11:30	14:28
Time finish sampling	11:40	13:15	11:26	14:02	14:55	11:22	12:50	15:37
Total time aerial exposure (min)	91	188	101	58	111	89	80	69

**Table S4.** Deployment details (times UTC) performed on behavioural discards *Nephrops*.

	Prawn	1	2	3	4	5	6	7	8	9	10
<b>Deployment 1</b>	Trawl began	06:00									
	Trawl tow ended	07:30									
	Prawns transferred to Calanus	07:45									
	Prawns placed under camera	09:09									
	Post-trawl treatment	~90 mins in air									
	Air temperature (°C)	9.5									
	Vigour	Individual vigour not recorded but a mix of C and Moribund									
	Comments	Test dive (sizes of animals not recorded)									
<b>Deployment 2</b>	Trawl began	07:45									
	Trawl tow ended	09:45									
	Prawns transferred to Calanus	10:00									
	Prawns placed under camera	10:50									
	Post-trawl treatment	~ 60 mins in air									
	Air temperature (°C)	16.5									
	Size (mm)	28	30	30	28	28	30	30	30	31	29
	Vigour	B	A/B	A	A/B	B	B	B	A/B	B	B
Comments	Spare animals were sunk in tubes at 20 m and also used in deployment 3										
<b>Deployment 3</b>	Trawl began	As above									
	Trawl tow ended	As above									
	Prawns transferred to Calanus	As above									
	Prawns placed under camera	11:20									
	Post-trawl treatment	~ 30 mins air, 65 mins recovery at 20 m									
	Air temperature (°C)	17.0									
	Size (mm)	25	27	28	28	29	27	28	28	31	32
	Vigour	B	B	B	A	A/B	C	B/C	B	B	B
Comments											
<b>Deployment 4</b>	Trawl began	10:30									
	Trawl tow ended	12:30									
	Prawns transferred to Calanus	12:43									
	Prawns placed under camera	13:04									
	Post-trawl treatment	~ 30 min air									
	Air temperature (°C)	17.3									
	Size (mm)	28	26	25	28	26	26	29	26	27	27
	Vigour	A	B/C	B/C	B	B	A/B	B	B	B	B
Comments											
<b>Deployment 5</b>	Trawl began	13:00									
	Trawl tow ended	14:40									
	Prawns transferred to Calanus	15:00									
	Prawns placed under camera	15:30									
	Post-trawl treatment	~ 50 mins air									
	Air temperature (°C)	17.3									
	Size (mm)	25	25	25	25	25	28	26	29	26	22
	Vigour	B/C	B/C*	B/C							
Comments	Deployment used to diver observations and to take flash stills photographs, video not analysed further										

**Table S4.** Deployment details (times UTC) performed on behavioural discards *Nephrops*.**Supplementary Table 1.** Summary of video clips.

Filename	Timestamp (hh:min:sec)	Time since release (min:sec)	Summary main events
Montage_Deploy1	09:09:23	00:00	Prawns released into arena
	09:09:44	00:21	2 prawns show almost immediate recovery
	09:09:48	00:25	Diver withdraws from site
	09:10:11	00:48	Prawn at top of group exhibits a flick-tail escape response, another prawn begins to recover and move around
	09:10:53	01:30	Another prawn recovers and moves away; 5 unrecovered animals remain
	09:12:00	02:37	Four more prawns begin to show signs of recovery, limb movement, attempts at righting
	09:14:53	05:30	All prawns except one showing some signs of recovery although three animals have still failed to fully right themselves.
	09:16:23	07:00	All prawns except two have either left the arena or are moving within the arena
	09:18:13	08:50	Prawn at top left interacts with another recovering animals and exhibits a flick-tail escape
	09:18:53	09:30	All prawns righted, all except one moving around, squat lobster enters arena and one prawn evades
	09:19:33	10:10	Prawn nips squat lobster which moves away, one prawn still only partially righted (camera 5)
	09:20:07	10:44	Squat lobster attacks partially righted prawn
	09:20:36	11:13	First crab arrives and chases off squat lobster
	09:20:39	11:16	Crab attacks recovered prawn but prawn exhibits escape response
	09:21:13	11:50	Crab attacks damaged prawn which exhibits a flick-tail response but then fails to right itself (moving from camera 5 to 6)
	09:21:31	12:08	Second crab then moves in to attack a separate prawn (camera 5)
09:21:37	12:14	Both crabs then attack the prawn	
09:21:53	12:30	Prawn is dragged out of camera arena by one crab and second crab resumes searching	
Montage_Deploy2	10:49:54	00:00	Prawns placed in camera arena
	10:50:01	00:07	Eight or nine prawns seem to recover almost immediately and begin to disperse
	10:50:08	00:14	Remaining prawns right themselves and

**Table S4.** Deployment details (times UTC) performed on behavioural discards *Nephrops*.

	10:51:01	01:07	begin moving around
			All animals appear to be moving around within, or have exited the arena, rate of tidal flow picking up compared with earlier deployment
	10:51:44	01:50	Single prawn re-enters arena, moving around on seabed
	10:51:50	01:56	Second prawn re-enters after exhibiting tail-flick escape response from some unseen stimulus
	10:53:05	03:11	End
Montage_Deploy3	11:20:54	00:00	Tide had picked up and diver reported that the prawns were all active but dispersed very rapidly, video of a single animal walking within the arena
	11:21:35	00:39	End
Montage_Deploy4	13:04:35	00:00	Prawns placed in arena, camera 3 not working
	13:04:40	00:05	Four animals show good recovery and move away
	13:04:45	00:10	All animals have righted themselves and are beginning to recover
	13:05:02	00:27	Three animals remain within arena
	13:05:44	01:09	Four prawns continue moving around within arena
	13:06:31	01:56	Two prawns show interaction then move apart
	13:06:49	02:14	
	13:10:00	05:25	End
Montage_Deploy5			Not analysed, divers using flash still photography

0

1 **Suppl. Fig. 1**

A)



2

3 **B)**



Suppl. Fig. 2.

